First phase of macroinvertebrate repopulation of Lake Orta (Buccione Basin) after liming

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ABSTRACT

The research described here was performed during 1993 and 1994 in the context of studies on the recovery of Lake Orta from acidification after the liming conducted from May 1989 to June 1990. The causes of the pollution were due to the oxidation of ammonium sulphate discharged from a factory producing rayon by a cupro-ammoniacal method and to the loads of heavy metals (Cu, Cr, Ni, Zn) derived from plating factories. One year after liming, an increase in pH and a marked decrease in metal concentrations were observed, and as a consequence some qualitative and quantitative changes were observed in 1992-93 in the planctonic communities and in the littoral macrobenthic fauna. This paper presents qualitative and quantitative data on macrobenthos collected along a transect between the littoral and the maximal depth in the Buccione Basin, which was directly affected by the effluents of the rayon factory. The data revealed some quantitative repopulation of this community, especially along the littoral. Qualitatively, however, the macrobenthos was still quite poor, consisting essentially of Oligochaeta and Diptera Chironomidae. The former were rare and mainly represented by Naididae between 15 and 22 m, while the Tubificidae were well represented in the littoral and the profundal zones. As regards the Chironomidae, the genera Procladius and Chironomus gr. thummi extended as far as the maximum depth, while others, such as Ablabesmyia, Parakiefferiella bathophila and Dicrotendipes gr. nervosus were mostly distributed in the littoral layers. Among the other taxonomic groups only the Trichoptera and the Diptera Ceratopogonidae were quantitatively significant. However, many systematic entities typical of the less polluted Italian subalpine lakes are still absent. In fact, the colonisation of Lake Orta by macrobenthic organisms is still influenced by the increased concentrations of heavy metals in the sediment precipitated from the water column. The research included analysis of lake water and sediments chemistry to obtain information on the trophic level of the environment.

Key words: Orta, acidification, liming, macroinvertebrate community

1. INTRODUCTION

1.1. Historical information

The first data we have on the benthic fauna of Lake Orta were provided by Parona (1880), in his report on a survey made by Pavesi (1879). The presence of some species of Tardigrada and Nematoda was recorded, along with that of Oligochaeta (Naididae and Aelosomatidae). Later on, in 1924, Colosi found the turbellarian Dendrocoelum lacteum to a depth of 120 m. During the biological decline of the lake, which began in 1927, macro- and microbenthic organisms were primarily observed along the zone of the lake nearest the shoreline. Pini (quoted by Monti, 1930) recorded finding two species of Lamellibranchia (Anodonta palustris and Unio cusianus), and Baldi (1949) observed in the Buccione Basin larvae of Diptera Chironomidae along with Nematoda, and abundant Protozoa populations (Difflugia and Centropyxis).

An important contribution to our knowledge of the benthic fauna of the lake was made by Moretti (1954a, b), who studied the distribution of Trichoptera along the shore of the whole lake, in relation to some environmental parameters which were of fundamental importance in determining the presence of the various species. He found 13 species of Trichoptera in the top 2 m; among them, only *Mystacides azurea* was numerically

well-represented in most of the 28 stations selected. *Ec-nomus tenellus* and *Oecetis* sp. were widely present but less abundant.

In addition to Trichoptera, other orders of Insecta were found, particularly Diptera Chironomidae, but also Hemiptera, Plecoptera, Odonata, Coleoptera, Megaloptera. There were also sporadic sightings of Porifera, Turbellaria Tricladida, Oligochaeta (Tubificidae), Protozoa and Copepoda.

The presence in large numbers of Trichoptera and Diptera Chironomidae in the benthic littoral community was confirmed by Oioli (1969), who also observed significant populations of Oligochaeta (especially Naididae). In addition, the presence of the following was recorded: Nemertea, Nematoda, Turbellaria, Irudinea, Hydracarina, Anphipoda, and some orders of Insecta, such as Collembola, Ephemeroptera, Plecoptera, Odonata, Coleoptera, Diptera (Ceratopogonidae, Limoniidae, Tipulidae).

Before the seventies, not much information was available on the benthic community in the deep zone of the lake. It was either found to be absent (Corbella *et al.* 1958), or essentially made up of Protozoa, Rotifera and Copepoda (Ruggiu 1969). However, in the winter (January-February) and summer (July) of 1971, some macroinvertebrates (Anphipoda, Diptera Chironomidae, Oligochaeta) were found between depths of 3 and 62 m at the mouths of some watercourses (Bonacina et al. 1988).

There were important changes in the deep benthic fauna at the beginning of the eighties, when a treatment plant was set up to recover ammoniacal nitrogen and to further reduce the copper loading, the reduction of which had begun in 1958. At the same time a plant for treating domestic and industrial sewage was activated on the western shore of the lake (Lacqua *et al.* 1983; Bonacina & Bonomi 1985). Bonacina & Bonomi (1984) had already recorded the presence of dense Oligochaeta populations, belonging to the species *Tubifex tubifex*, in the deep zone (between 20 and 140 m) of the three lake basins.

In the same period, less important changes occurred in the littoral macrofauna, which in 1987-88 was still largely made up of Trichoptera and Diptera Chironomidae. Other taxonomic groups (Irudinea and Oligochaeta Tubificidae) were however more widespread round the lake shore, at least in the top 150 cm (Ballarè *et al.* 1992). In this period Chironomidae and Trichoptera were also well represented in the periphyton (Cattaneo 1992).

Research on the littoral macrofauna continued between 1990 and 1993, both during the liming operation and immediately following it. The results revealed that the positive effects of the recovery intervention observed on the chemistry of the lake, produced no significative changes in the benthic macrofauna, particularly from a qualitative point of view. However, there had been a certain numerical increase of some taxonomic groups already present in the community in the years prior to the recovery operation (Tesauro *et al.* 1995). A study of the littoral macrobenthos performed in 1998 (Bielli & Tesauro 2001) essentially confirmed the data obtained in the previous studies.

After the liming intervention, two research projects were set up to verify the situation of the macrobenthic fauna colonising the deeper layers. The first performed in the Buccione Basin (De Margaritis 1995) and the second one extended to the whole lake basin (Sabolla 1998). The results showed that some groups of organisms, particularly Oligochaeta and Chironomidae, were well represented also in some areas of the sub-littoral and deep zones, where, however, the number of systematic entities forming these populations was still extremely small. Important information was also gained on the environmental quality of the sediments, which appear to be still partly affected by the toxic residues that had accumulated on the bottom as far back as the first stage of the pollution of the lake, but also since the liming intervention (Baudo et al. 1989, Baudo & Beltrami 2001).

1.2. Aims of the study

With the 1993-94 research on the macrofauna of the Buccione Basin, the most southerly of the basins of

Lake Orta, it was intended to analyse the effects of liming on the structure of the populations established at different levels between the littoral and the deep zones. This basin, with a maximum depth of 35 m and sloping gently from the shore to the bottom, was particularly suitable for the purpose.

We decided to focus on this basin because this part of the lake was highly polluted by the discharge from the Bemberg factory, and via the River Lagna, from the numerous plating factories in the area. This impact was the cause of the major changes in the abiotic and biotic components at sediment level (Corbella et al. 1958; Bonacina et al. 1986). The liming, which was started in this basin, resulted in an immediate improvement in the water chemistry. In particular, an increase in pH and alkalinity was recorded, as well as a marked reduction in concentrations of heavy metals, especially copper and aluminium, and of ammonia nitrogen, encouraging the appearance of new taxonomic entities at plankton community level (Ambrosetti et al. 1992; Calderoni & de Bernardi 1992). It was also important to verify the recovery times of the benthic macrofauna in this part of the lake, so that our knowledge of the state of this community could be subsequently extended to the other sectors of the lake.

Along with the study of the macrobenthos in the Buccione Basin, lake water and sediments chemistry were analysed to gain information on the quality and level of productivity of the environment after the recovery intervention.

2. METHODS

2.1. Water parameters

Lake water samples which was performed for the period 1993-94 for macrobenthic study continued in 1995. The samples were collected by Ruttner bottle almost monthly between the beginning of spring (March-April) and autumn (October-November) at station A (Fig. 1), at the point of maximum depth of the basin. On each date samples were taken at regular intervals (every 2 m) in the layers between 0-10 and 10-20 m; the samples from each layer were mixed to give an integrated sample. In addition, a discrete sample was taken at the depth of 30 m, close to the sediments. On every sampling date water transparency was also measured by Secchi disk.

Concentrations of nitrogen and total phosphorus, suspended particulate matter (total seston) and its components (organic carbon and nitrogen), and chlorophyll were measured on each sample.

Total nitrogen and phosphorus values were determined on non-filtered samples by spectrophotometry (Tartari & Mosello 1997). The water samples used for estimating total seston and chlorophyll were first filtered through a plankton net of 126 μ m mesh size. The material of a size between 1 and 126 μ m was then concentrated by vacuum filtering on Whatman GF/C fibre-

glass filters, previously combusted at 500 °C. Total seston was measured gravimetrically, taking account of the weight of the filter before and after filtration, and the volume of water filtered. The filters were then analysed using a CHN elemental analyser to evaluate the content of organic carbon (POC) and nitrogen (PON) in the particulate matter (Bertoni & Callieri 1982). The carbon/nitrogen ratio (C/N) was calculated on the data of the individual samples to find out the prevalent composition of the organic fraction of the seston. The inorganic fraction was estimated by gravimetry, subtracting from the total seston concentrations the POC values multiplied by a conversion factor (Lenz 1977). The chlorophyll concentrated on the filters was extracted with methanol and its content measured by spectrofluorimetry (Holm-Hansen & Rieman 1978).

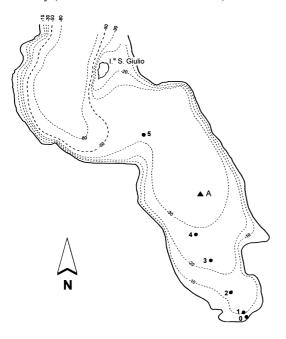


Fig. 1. Southern portion of Lake Orta. Location of the stations for sampling water (A) and sediment $(0\div 5)$.

2.2. Sediment and macrobenthic community parameters

The samples for studying the sediment and the macrobenthic fauna were taken at the same time in 6 stations (0÷5), respectively at depths of 4, 7, 15, 22, 28 and 35 m, distributed along a northward transect between the shoreline and the island of S. Giulio (Fig. 1). Sampling was done seasonally in 1993 (March, July and November) and 1994 (March, July and October) by Ponar dredge (collection area: 420.25 cm²). Three samples were taken at each station and on each date: one for sediment analysis and two for the study of the macrobenthic fauna.

The samples to be used for sediment analysis were kept at -20 °C until analysed; the macrobenthic samples were sieved through a 270 μ m mesh nylon net and fixed with enough neutral formaldehyde to prevent damage to the animals.

Each sediment sample was analysed for its concentrations of total organic matter, and for some parameters (organic and inorganic carbon and nitrogen, and plant pigments) which could give information on its origin and qualitative composition. It must be remembered that the organic matter in sediments is made up both of allochthonous elements from the surrounding land, and autochthonous elements from the pelagic layers, arriving via sedimentation, and from the littoral macrophyte vegetation. It is also the main food source of the macrobenthic populations, particularly of the detritivorous populations which are most closely associated with the sediment. Its qualitative and quantitative variations in time and space are thus a major factor in the settlement and growth of the populations. The calcium carbonate content was also estimated to verify whether the massive inputs of the element into the water column during the liming operation had resulted in its accumulation on the lake bottom. Each parameter was analysed on fractions of whole sediment taken from each sample. The analyses were done in triplicate and the values resulting for each parameter averaged.

The organic matter values were estimated as loss on ignition (LOI); the resulting weight difference was measured by heating to a temperature of 550 °C a sample previously dried at 105 °C and weighed. The calcium carbonate concentrations were calculated by taking the same samples from a temperature of 550 °C to 950 °C and multiplying the resulting weight difference by the factor of 2.27, which takes into account the loss of CO₂ during combustion at high temperature (Wetzel 1970; Lewis & Weiberhan 1981). The concentrations of organic matter and of carbonates are expressed as a percentage of the dry weight.

The content of organic carbon and nitrogen was evaluated by CHN elemental analyser on known quantities of sediment dried at 105 °C, and that of inorganic carbon and nitrogen on samples combusted at 550 °C. The resulting values are expressed as μg per mg of sediment, and as results from a simple proportion, the percent incidence of these parameters in the sample may be found. To evaluate the prevalent origin composition of the organic matter, the carbon/nitrogen ratio (C/N)

		TN (mg l ⁻¹)			TP (μg l ⁻¹)	
	0-10 m	10-20 m	30 m	0-10 m	10-20 m	30 m
1993						
25-Mar	-	-	-	-	-	-
03-May	3.54	3.48	4.12	8.22	7.80	129.11
20-May	3.40	3.47	3.46	7.80	5.71	5.71
21-Giu	3.20	3.35	3.35	5.29	6.13	5.71
26-Jul	2.93	3.18	3.24	7.81	4.87	4.87
31-Aug	3.40	3.58	3.73	4.47	4.47	5.29
30-Sep	3.24	3.32	3.51	5.29	6.13	5.29
25-Oct	2.82	2.87	3.29	5.29	6.13	4.87
04-Nov	2.85	2.98	3.40	5.29	5.29	4.45
17-Nov	2.87	2.90	3.29	5.71	5.71	4.45
Annual mean	3.14	3.24	3.49	6.13	5.80	18.86
1994						
14-Mar	3.02	2.99	3.02	4.45	4.87	4.87
20-Apr	3.07	3.07	3.02	5.29	5.29	5.29
30-May	3.04	3.13	3.17	-	-	-
21-Giu	2.94	2.93	3.05	5.84	5.84	5.84
19-Jul	2.91	3.08	3.17	3.34	5.84	6.26
18-Aug	2.68	2.77	2.90	4.59	4.59	5.01
19-Sep	2.71	2.79	2.97	5.01	5.42	4.59
26-Oct	2.77	2.81	3.09	3.34	3.34	2.92
Annual mean	2.89	2.95	3.05	4.55	5.03	4.97
1995						
03-Apr	2.87	2.95	2.85	4.59	4.17	4.17
26-Apr	2.89	2.84	2.87	6.26	6.67	6.26
24-May	2.78	2.81	2.89	5.42	6.68	6.26
21-Giu	2.72	2.72	2.83	4.59	4.59	4.59
19-Jul	2.60	2.80	2.86	5.42	5.42	4.59
24-Aug	2.45	2.65	2.69	2.50	4.17	4.17
21-Sep	2.52	2.65	2.76	3.76	5.84	3.76
18-Oct	2.56	2.64	2.81	3.76	2.92	4.59
Annual mean	2.67	2.76	2.82	4.54	5.06	4.80

Tab. 1. Monthly concentrations and annual means of total nitrogen (TN) and phosphorus (TP) in years 1993÷1995.

On each sample reserved for macrobenthos study, the organisms were sorted from the sediment, the individuals preserved in bottles containing neutral 10% formaldehyde, and then identified to species or other systematic entity level and counted to evaluate their numerical density. The individuals from each pair of samples were counted together. After being counted, the organisms were washed in water, dried on filter paper and weighed on a Mettler balance to estimate their biomass. Before each weighing, the larvae and pupae of the Trichoptera and the Chironomidae were removed from their cases. It should be stressed that the weight of organisms preserved in formaldehyde tends to be lower than their original weight (Donald & Paterson 1977; Landhal & Nagel 1978). The biomass was therefore calculated as wet weight in formaldehyde (Wingberg 1971). Numerical density is expressed as ind m⁻², biomass as mg m^{-2} .

3. RESULTS

3.1. Water parameters

3.1.1. Total nitrogen and phosphorus

The mean total nitrogen concentrations along the water column showed a gradual decrease between 1993 and 1995 (Tab. 1).

In the top 20 m, the values went from 3.2 to 2.8 mg N Γ^1 . Slightly higher values (up to 3.05 mg N Γ^1) were found in 1993 at the depth of 30 m, where the samples were taken about 50 cm from the bottom. Over the three years of observation, seasonal variations were small along the whole column; the values were slightly higher between spring and summer in the top 20 m layer. The situation found in the Buccione Basin is reflected by the trend of the mean total nitrogen concentrations in the whole lake basin. Variations in time are closely related with those of nitrate, which after the liming and the dramatic reduction of ammonia nitrogen, is the most important component of the nitrogen cycle in the lake (Calderoni *et al.* 1997).

As regards total phosphorus concentrations along the column there was a slight decrease (Tab. 1). In 1993, mean total phosphorus content was between 5 and 6 μ g P Γ^1 , with values over 4 μ g P Γ^1 for most of the year and around 8 μ g P Γ^1 in the top 20 m in spring. The high value (129 μ g P Γ^1) recorded at 30 m in May has to be regarded as an isolated occurrence, probably caused by the release of phosphorus from the sediment following the decomposition of organic material accumulating on the bottom (Kelly 1991). In 1994 and 1995 the mean phosphorus content along the column fluctuated be-

tween 3.3 and 5 μ g P l⁻¹. However, higher values were found in spring-summer 1994 and again in spring 1995.

Interestingly, no great variations in the mean total phosphorus content compared to the years prior to the liming have been recorded. In fact, it is likely that the increase in pH determined by the changes which have occurred in the lake's acid-base system has caused the precipitation of phosphorus in the sediment in the form of insoluble metal hydroxides, in particular of iron and aluminium (Calderoni & Tartari 2001).

3.1.2. Suspended particulate matter

In the three years of this research, the mean concentration values of suspended particulate matter in the size range from 1 to 126 μ m were between 0.9 and 1.2 mg l⁻¹ in the top 20 m, and between 0.5 and 0.6 at 30 m (Tab. 2). In the top 20 m the total seston content showed marked variations during the year, with values up to 1.5 mg l⁻¹ in spring (April-May) and autumn (September-October), and higher values in summer (June-August). The highest concentration (3.7 mg l⁻¹) was measured in July 1993 during an algal bloom of the chlorophyceans *Cosmarium tenue* and *Stichococcus minutissimus*. At 30 m the total seston content showed lower seasonal variations, with the highest values (up to 0.9 mg l⁻¹) in spring.

Tab. 2. Annual means recorded in years 1993÷1995 for total seston, the inorganic and organic fraction, and the content in organic carbon and nitrogen, and in chlorophyll.

	Depth	Aı	nnual me	an
	(m)	1993	1994	1995
Total seston (mg l ⁻¹)	0-10	1.120	1.163	0.894
	10-20	0.928	0.973	1.019
	30	0.607	0.594	0.616
Inorganic seston (mg l ⁻¹)	0-10	0.344	0.377	0.401
	10-20	0.350	0.302	0.517
	30	0.229	0.186	0.302
Organic seston (mg l ⁻¹)	0-10	0.776	0.786	0.493
	10-20	0.578	0.671	0.502
	30	0.378	0.407	0.314
Organic carbon (µg l ⁻¹)	0-10	388	393	247
	10-20	289	336	251
	30	189	204	157
Organic nitrogen ($\mu g l^{-1}$)	0-10	42	41	32
	10-20	33	34	36
	30	22	23	23
C/N Ratio	0-10	9	10	8
	10-20	9	10	7
	30	9	9	7
Chlorophyll (µg l ⁻¹)	0-10	4.2	3.3	2.0
	10-20	3.5	5.0	3.2
	30	1.9	2.5	1.5

The distribution in space and time of POC and PON, and chlorophyll, essentially reflects that of total seston (Tab. 2). On average, the highest content of these pa-

rameters was found in the top 20 m, with the highest values in spring-summer and autumn. Fairly high quantities (POC = 1213 μ g Γ^1 ; PON = 125 μ g Γ^1 ; Chl = 15.5 μ g Γ^1) were measured in July 1993, at the same time as the maximum observed for total seston, and in October 1994 (POC = 684 μ g Γ^1 ; PON = 92 μ g Γ^1 ; Chl = 7.5 μ g Γ^1), during a bloom of chlorophyceans and diatoms. However, mean concentrations of a certain importance were also measured at 30 m, where living particulate matter is probably still well represented along with the matter deposited in the sediment from the water column.

An interesting observation is that the annual fluctuations of seston and its components (POC, PON and chlorophyll) essentially reflect the succession and the numerical participation during the year of the main taxonomic groups in the phytoplankton community, revealed by the seasonal variations in water transparency as measured by Secchi disk (Fig. 2). Important among the phytoplankton, both as regards density and biomass, are the chlorophyceans (present for most of the year), accompanied by crisophyceans in spring and by cyanophyceans and diatoms in summer-autumn (Morabito & Curradi 1997; Morabito *et al.* 2001).

Elements of autochthonous origin, phytoplankton in particular, have undoubtedly made up a quantitatively significant fraction of the suspended particulate matter in this lake. This is suggested primarily by the relationships existing between the contents of POC, chlorophyll, and the organic fraction (Tab. 2). In 1993 and 1994 the concentrations of these elements maintained values which were on average higher than those found in 1995. Moreover, in the first two years the organic fraction represented on average between 60 and 70% of total seston, with contents above 60% for most of the year (Fig. 3). In contrast, PON concentrations along the water column remained at values which were more or less similar for the three years of the study.

The role of phytoplankton component in the organic fraction of the seston are also suggested by the values of the C/N ratio measured along the column, which are on average between 9 and 10 in 1993 and 1994, and between 7 and 8 in 1995. The 1994 values of this ratio showed a seasonal variability higher than that observed in 1993 and 1995, reaching a maximum of 18 in August, when high values of total seston (1433 mg Γ^1), POC (562 µg Γ^1) and chlorophyll (3.7 µg Γ^1) were recorded, but low values of nitrogen (32 µg Γ^1).

The data collected during this study provide a description of the trophic situation of the lake in the first years after the liming intervention. The changes in the water chemistry have not actually had a profound effect on the trophic state of the lake, which has remained at a medium-low level. The values found for total seston and its components are not in fact very different from those measured in Lake Maggiore at the beginning of the 90s, when the lake was in a meso-oligotrophic condition (Bertoni & Callieri 1993).

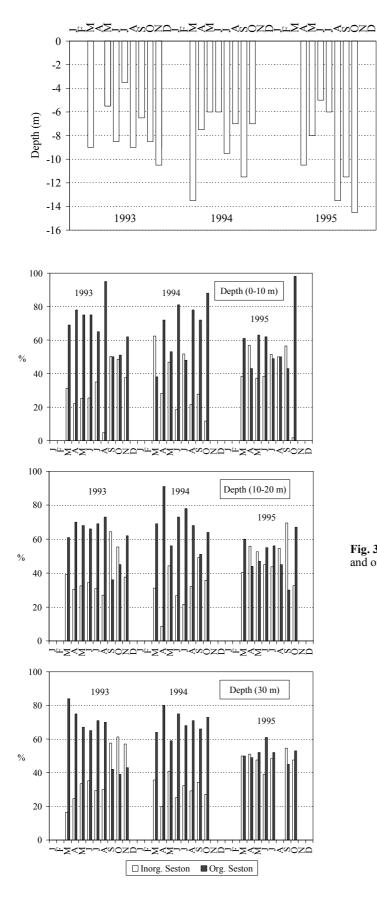


Fig. 2. Secchi disk transparency for period 1993÷1995.

Fig. 3. Monthly variations of percent values of inorganic and organic seston recorded at different depths.

Tab. 3. Seasonal concentrations and annual means of organic carbon and nitrogen, expressed in μ g mg⁻¹ of dry weight of sediment, and respective values of C/N ratio. Seasonal values and annual means, expressed as unit per gram of organic matter, of chlorophyll derivatives (SPDU₆₆₅), of total carotenoids (SPDU₄₅₀), and of respective ratio chlorophyll derivatives/total carotenoids (CD/TC).

		1993						1994						
		C org.	N org.	C/N	SPDU ₆₆₅	SPDU ₄₅₀	CD/TC	C org.	N org.	C/N	SPDU ₆₆₅	SPDU ₄₅₀	CD/TC	
St. 0	March	-	-	-	-	-	-	33.40	2.10	15.90	99.25	211.04	0.47	
(4m)	July	33.30	2.20	14.90	68.46	200.29	0.34	29.20	2.60	11.40	65.85	163.75	0.40	
	November	37.80	2.80	13.30	133.50	255.94	0.52	33.30	2.80	11.90	69.35	161.77	0.43	
	Annual mean	35.55	2.50	14.10	100.98	228.11	0.43	32.00	2.50	13.10	78.15	178.85	0.43	
St. 1	March	79.70	6.80	11.70	199.05	331.08	0.60	33.20	2.60	12.70	192.27	302.59	0.64	
(7m)	July	29.50	2.20	13.20	140.97	240.70	0.59	29.40	2.60	11.20	158.43	268.01	0.59	
	November	37.30	3.40	11.10	147.06	241.74	0.61	28.50	2.50	11.50	182.43	269.45	0.68	
	Annual mean	48.83	4.13	12.00	162.36	271.17	0.60	30.40	2.60	11.80	177.71	280.02	0.63	
St. 2	March	63.10	6.00	10.50	190.00	369.01	0.52	55.00	4.60	12.00	169.50	333.48	0.51	
(15m)	July	67.70	6.50	10.40	200.73	374.44	0.54	50.80	5.20	9.80	166.72	339.80	0.49	
` <i>´</i>	November	47.00	4.70	10.10	124.76	264.88	0.47	59.40	6.50	9.20	195.08	358.77	0.54	
	Annual mean	59.27	5.73	10.33	171.83	336.11	0.51	55.07	5.40	10.30	177.10	344.02	0.51	
St. 3	March	50.60	5.10	10.00	172.68	374.75	0.46	54.50	4.50	12.10	81.44	192.20	0.42	
(22m)	July	62.10	5.80	10.70	130.19	289.73	0.45	44.20	5.00	8.90	74.02	181.52	0.41	
	November	57.60	5.90	9.80	116.59	256.28	0.46	45.40	5.10	8.80	74.78	182.58	0.41	
	Annual mean	56.77	5.60	10.17	139.82	306.92	0.46	48.03	4.90	9.90	76.75	185.43	0.41	
St. 4	March	64.20	6.50	9.90	77.00	208.21	0.37	-	-	-	38.28	121.19	0.32	
(28m)	July	49.60	4.50	10.90	68.11	196.78	0.35	47.30	5.10	9.30	75.95	188.01	0.42	
	November	64.20	6.10	10.50	78.67	204.53	0.39	46.40	5.20	8.90	53.24	154.45	0.35	
	Annual mean	59.33	5.70	10.43	74.60	203.17	0.37	46.90	5.20	9.10	55.82	154.55	0.36	
St. 5	March	53.30	5.00	10.60	87.22	255.92	0.34	47.00	4.50	10.40	44.66	139.21	0.32	
(35m)	July	47.00	4.60	10.30	47.80	142.58	0.34	49.50	5.00	9.80	83.23	256.87	0.32	
` /	November	46.30	4.40	10.50	22.53	126.16	0.18	42.80	4.30	10.10	38.12	128.01	0.30	
	Annual mean	48.87	4.67	10.47	52.52	174.89	0.29	46.40	4.60	10.10	55.34	174.70	0.31	

The rather low trophic state of the lake between 1993 and 1995 is largely due to the quite low phosphorus concentrations in water after the liming (Calderoni & Tartari 2001). However, the lake's trophic state is expected to increase in the future with the evolution of the water chemistry, and the high phosphorus loads which continue to arrive from the catchment area, especially from diffused sources (Calderoni *et al.* 1997).

3.2. Parameters of the sediments

There were no major variations in the distribution in space of the parameters considered over the two years of the study (1993-94).

3.2.1. Calcium carbonate

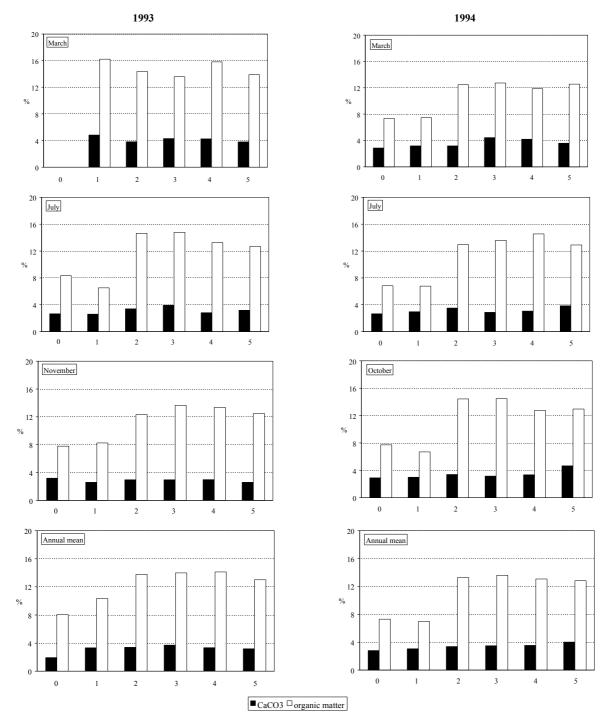
The calcium carbonate concentrations were generally not particularly high, and reflect the lithological composition of the watershed, which is made up mostly of siliceous rocks (Boriani & Sacchi 1974). During the two years of the study, calcium carbonate contents on average between 2.9 and 4% were estimated, with slight variations along the bathymetric profile, where values higher than 4% were estimated especially in spring and autumn (Fig. 4).

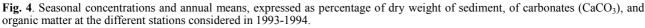
The observed values show clearly that, at least in the early stages following the liming, there was no significant increase in carbonate concentrations at sediment level. This confirms findings in other limed lake environments (Driscoll *et al.* 1989). As regards Lake Orta, it was calculated that around 85% of the calcium carbonate added to the lake dissolved during the early stages of the intervention, with a consequent increase in concentrations of calcium ion in the water column which reached values of around 13 mg l^{-1} in 1993-94 (Calderoni *et al.* 1993).

3.2.2. Organic matter, organic carbon and nitrogen and plant pigments

The highest mean concentrations of organic matter (between 13 and 14%), organic carbon (between 5 and 6%) and organic nitrogen (between 0.5 and 0.6%) were observed between 15 and 35 m. Variable values were found during the year between 15 and 28 m, where in 1993 the contents of organic matter and carbon were on average higher than those of 1994 (Fig. 4 and Tab. 3). In contrast, organic nitrogen concentrations were less variable in space and time in the whole of the sub-littoral and profound layer (between 15 and 35 m).

In spring 1993, high contents of organic matter (16.2%), organic carbon (8%) and nitrogen (0.7%) were also estimated in the littoral layer (7 m); however, during the year, lower concentrations were recorded in this layer than in the underlying layers. Interestingly, the seasonal variations in the contents of these parameters along the bathymetric profile are linked to the different





contribution of the organic matter reaching the sediment. In fact, the concentrations of inorganic carbon and nitrogen along the whole bathymetric profile were extremely low during the whole year, with values of zero or close to zero.

Between 15 and 35 m the autochthonous component, especially that of plant origin, was well represented in the organic matter of the sediment. The C/N ratio pre-

sented values on average between 9 and 10, with little variation during the year. Higher values of the ratio (between 12 and 14) were estimated in the littoral layers (4 and 7 m), where, along with the autochthonous component, there is probably a significant contribution of allochthonous matter from the shore environment, especially at certain times of year (Wetzel 1975; Håkanson & Jansson 1983).

Information on the qualitative composition of the sedimented organic matter is also provided by the distribution in space of plant pigment concentrations; these show significant seasonal variations with depth, reflecting the temporal dynamics of their deposition and degradation at sediment level.

Both in 1993 and 1994, higher mean contents were recorded between depths of 7 and 15 m, with a maximum of 178 for chlorophyll derivatives and 344 for carotenoids. Lower values, especially for the chlorophyll derivatives, were recorded between 28 and 35 m (Tab. 3). In this depth interval values of the CD/CT ratio were measured between 0.2 and 0.4, confirming the importance of the autochthonous component, especially of phytoplankton, in the organic matter of the sediment (Sanger & Gorham 1970; Wetzel 1970; Gorham & Sanger 1975; Swain 1985).

The presence of plant pigments of largely autochthonous origin was also revealed by paleo-limnological studies carried out in the deep zone of the lake by Guilizzoni & Lami (1990) in the years prior to the liming. The accumulation of these pigments in the sediment was probably caused by an increase in algal production following the removal of the ammoniacal nitrogen and copper loads by the Bemberg factory from the 80s, and also to the high acidity of the water, which facilitated the preservation of the pigments at bottom level. Nevertheless, the high carotenoid content seems to suggest that the autochthonous component (algae and macrophytes) is also important in the sediment of the shallower layers (between 4 and 22 m), where higher values (on average between 0.4 and 0.6), of the CD/CT ratio were recorded. However, even higher values (between 0.6 and 0.7) were measured at the depth of 7 m, where a greater quantity of allochthonous material of plant origin arrives from the shore.

3.3. Macrobenthic fauna

In the two years of study the stations along the transect considered showed a qualitatively rather poor macrobenthic community composed mainly of Oligochaeta and Diptera Chironomidae. The other groups of organisms found, such as Turbellaria, Hydracarina, and some Insecta orders (Diptera Ceratopogonidae, Trichoptera, Megaloptera), were poorly represented. The most numerous populations were observed between 4 and 22 m, where mean densities varying from 1030 to 6258 ind m² were recorded (Tabs 4 and 5).

The Oligochaeta (Figs 5 and 6), comprising some species of Naididae and Tubificidae, were present in higher densities in the littoral layers (4 and 7 m), with values varying over the year from 400 to 2177 ind m^{-2} , correlated to the presence of young and mature individuals of *Limnodrilus hoffmeisteri*, which dominated the population numerically. The Oligochaeta were in contrast very scarce at 15 m, and in the underlying layers the quali- and quantitative structure of the popula-

tion showed some major differences in the two years of the study.

In 1993, not negligible densities of Naididae particularly of *Veidovskiella comata* e *Nais communis*, were found at 22 m, where the Tubificidae were absent. In contrast, the Tubificidae were present in significant numbers between 28 and 35 m, where they were the dominant element of the macrobenthos due to the numbers of young and mature individuals of *Tubifex tubifex*. In 1994 Oligochaeta were not recorded at 22 m, and in the underlying layers the density of *Tubifex tubifex*, the dominant element of the population, was lower than was recorded in 1993.

The Chironomidae were made up of immature individuals (larvae and pupae) belonging to the sub-families of the Tanypodinae, Orthocladiinae and Chironominae.

During the year their density was fairly high in the top 22 m, where they made up over 60% of the total density, due especially to the contribution of the genera *Procladius* e *Chironomus thummi* gr., both of which were present in low numbers at greater depths. *Parakiefferiella bathophila* e *Dicrotendipes nervosus* gr. had some quantitative importance in the littoral layers (4 and 7 m). In 1994, due to the marked decrease in numbers of the Oligochaeta, the Chironomidae were, along with the genus *Zavrelimyia*, the dominant element of the macrobenthos also at a depth of 28 m.

The other groups are of little importance in the total density of the benthos, with percentages between 1 and 9%. Numerically the most significant are the Trichoptera, particularly *Mystacides azurea* and the Ceratopogonidae, which are present mainly in the shallower layers. In 1994 the Turbellaria were of some importance between 22 and 28 m.

The trend of benthic biomass at the sampling stations generally reflects the trend of the densities (Tabs 4 and 5). High biomass values were observed in the top 22 m, due largely to the contribution of the Diptera Chironomidae, which on average represent over 50% of the total biomass. In this depth interval the biomass values of the Chironomidae vary over the years in relation to the biological cycle of the elements which are numerically best represented (*Procladius* e *Chironomus thummi* gr.). The biomass of the Chironomidae (Figs 5 and 6), which is high in spring, tends to decrease during summer and autumn as the mature larvae emerge and the young larvae of the new generation, smaller and with a low mean individual weight, increase.

In the littoral layers (4 and 7 m), the Oligochaeta too represent a significant component of the total biomass, largely due to the contribution of young and mature individuals of *Limnodrilus hoffmeisteri* and of *Branchiura sowerbyi*. At certain moments in the season *Branchiura sowerbyi*, which represents quite a small part of the population in numerical terms, may have a major impact on the benthic biomass because of the greater mean weight of the individuals.

Tab. 4. Annual mean values for density and biomass, recorded in 1993, for different taxonomic entities in the various stations.
Values relating to Chironomidae include larvae and pupae. Percentages are also presented.

			Density				Biomass (mg m ⁻²)						
Depth (m)	St. 0 (4m)	St. 1 (7m)	St. 2 (15m)	St. 3 (22m)	St. 4 (28m)	St. 5 (35m)	St. 0 (4m)	St. 1 (7m)	St. 2 (15m)	St. 3 (22m)	St. 4 (28m)	St. 5 (35m)	
OLIGOCHAETA													
NAIDIDAE													
Chaetogaster	-	-	-	24	8	4	-	-	-	5	1	1	
Pristina longiseta	60	-	-	-	-	-	1	-	-	-	-	-	
Vejdovskiella comata	-	-	-	250	52	-	-	-	-	10	2	-	
Nais communis	-	-	28	174	52	8	-	-	2	11	4	0	
Nais elinguis	-	-	8	-	-	-	-	-	0	-	-	-	
TUBIFICIDAE		-	-	-	-	-	-	-	-	-	-	-	
Bothrioneurum vejdovskianum	24	-	-	-	-	-	3	-	-	-	-	-	
Branchiura sowerbyi	18	56	-	-	-	-	21	2674	-	-	-	-	
Tubifex tubifex	-		-	-	222	28	-	_	-	-	1337	123	
immature Tubificidae	6	75	-	-	250	508	5	20	_	-	561	718	
Limnodrilus hoffmeisteri	2059	186	4	-	12	4	1105	70	1	-	6	44	
immature Limnodrilus	-	321	_	-	-	-	-	66	-	-	-	-	
Oligochaeta total	2167	638	40	448	596	552	1135	2830	3	26	1911	886	
%	35	21	3	28	85	79	33	48	0	1	82	79	
	55	21	5	28	85	19	55	40	0	1	62	13	
HYDRACARINA													
PIONIDAE	-	-	-	-	-	4	-	-	-	-	-	52	
Hydracarina total	-	-	-	-	-	4	-	-	-	-	-	52	
%	-	-	-	-	-	1	-	-	-	-	-	5	
INSECTA													
Diptera													
CHIRONOMIDAE													
Tanypodinae													
Procladius	1065	1376	547	210	8	24	541	1238	1074	489	26	28	
Macropelopia	-	-	-	4	4	4	-	-	-	14	3	4	
Ablabesmyia	149	87	99	-	-	4	12	11	30	-	-	5	
Zavrelimyia	6	12	135	143	32	91	12	2	33	- 95	17	41	
Orthocladiinae	0	12	155	145	52	91	1	2	55	95	1 /	41	
Paracladius	48	-	_				4	-	-				
Cricotopus (Isocladius)	36	-	-	-	-	-	3	-	-	-	-	-	
	-	- 4		-			-	0				-	
Cricotopus (Cricotopus)	- 54	4 297	-4		-	-	- 4	27	-0	-	-	-	
Parakiefferiella bathophila	54	297	4	-	-	-	4	27	0	-	-	-	
Chironominae		4						0					
Pagastiella orophila	-	4	-	-	-	-	-	0	-	-	-	-	
Phaenopsectra	137	8	-	4	4	-	19	0	-	4	5	-	
Chironomus thummi gr.	625	460	436	801	55	16	1234	1686	1174	4067	375	112	
Polypedilum nubeculosum gr.	78	36	-	-	-	-	23	18	-	-	-	-	
Dicrotendipes nervosus gr.	1678	75	-	-	-	-	465	16	-	-	-	-	
Demicryptochironomus	6	-	-	-	-	-	1	-	-	-	-	-	
Cladopelma lateralis gr.	-	8	-	-	-	-	-	1	-	-	-	-	
Chironomidae total	3882	2367	1221	1162	103	139	2307	2999	2311	4669	426	190	
%	62	77	97	72	15	20	66	51	90	99	18	17	
CERATOPOGONIDAE	54	16	-	-	-	-	11	5	-	-	-	-	
Megaloptera	-	-						-					
Sialis	-	-	4	-	-	-	-	-	259	-	-	-	
Trichoptera			-										
Ecnomus tenellus	-	8	-	-	-	-	-	3	-	-	-	-	
Mystacides azurea	155	48	-	-	-	-	27	28	-	-	-	-	
Other insecta total	209	72	4	_	-	-	38	36	259	-	_	-	
%	209	2	4	-	-	-	1	1	10	-	-	-	
												-	
GENERAL TOTAL	6258	3077	1265	1610	699	695	3480	5865	2573	4695	2337	1128	

The values of the Oligochaeta biomass were very low between 15 and 22 m, where they were very scarce or represented by Naididae, which are small in size. In contrast, in the underlying layers (28 and 35 m) *Tubifex tubifex* represented almost the entire macrobenthos, also in terms of biomass, especially in 1993. In 1994, at 28 m, the Chironomidae with the genus *Zavrelimyia* were dominant, although their density values were not high. The participation of the other groups of organisms in terms of biomass is not very different from their participation in terms of density, and represents between 1 and 10% of the macrobenthos. The commonest elements numerically play a significant role, especially *Mystacides azurea* among the Trichoptera and the Ceratopogonidae at 4 and 7 m, and the Turbellaria at 22 and 28 m.

			Density	(ind. m ⁻²)		Biomass (mg m ⁻²)						
Depth (m)	St. 0 (4m)	St. 1 (7m)	St. 2 (15m)	St. 3 (22m)	St. 4 (28m)	St. 5 (35m)	St. 0 (4m)	St. 1 (7m)	St. 2 (15m)	St. 3 (22m)	St. 4 (28m)	St. 5 (35m)
TURBELLARIA												
Tricladida												
Polycelis nigra	-	-	4	40	12	4	-	-	3	51	23	5
Turbellaria total	-	-	4	40	12	4	-	-	3	51	23	5
%	-	-	0	4	6	2	-	-	0	1	9	1
OLIGOCHAETA NAIDIDAE												
Dero digitata	-	-	4	-	4	_	-	-	1	-	2	_
Nais communis	-	-	4	-	-	_	-	-	0	-	-	_
Naididae others	_	4	4	_	_	_	_	0	0	-	_	-
TUBIFICIDAE	-	4	4	-	-	-	-	0	0	-	-	-
Branchiura sowerbyi	270	24	-	-	-	-	9104	37	-	-	-	-
Tubifex tubifex	-	-	-	-	4	-	-	-	-	-	10	-
immature Tubificidae	-	12	-	-	32	159	-	1	-	-	34	542
Limnodrilus hoffmeisteri	226	36	-	-	-	-	788	52	-	-	-	-
immature <i>Limnodrilus</i>	1015	825	12	-	_	-	634	358	3	-	-	_
Oligochaeta total	1511	901	24	-	40	159	10526	448	4	-	46	542
%	35	26	2	-	19	89	74	9	0	-	19	92
INSECTA	50	20	-		.,	07	, .	-	0			/ -
Diptera												
CHIRONOMIDAE												
Tanypodinae	017	064	760	100	4		505	1100	994	222	2	
Procladius	817	964	769	190	4 4	-	505	1108		333	2 30	-
Macropelopia	-	-	-	-		4	-	-	-	-		39
Ablabesmyia	8	43	4	12	-	-	3	15	0	1	-	-
Zavrelimyia	48	8	48	191	143	8	7	1	45	138	115	2
Orthocladiinae												
Psectrocladius sordidellus gr.	12	-	-	-	-	-	4	-	-	-	-	-
Cricotopus (Isocladius)	-	4	-	-	-	-	-	3	-	-	-	-
Parakiefferiella bathophila Chironominae	369	440	8	-	-	-	39	68	1	-	-	-
Paratendipes	4	-	-	-	-	-	0	-	-	-	-	-
Phaenopsectra	24	44	-	-	-	4	15	18	-	-	-	1
Chironomus thummi gr.	936	607	563	599	4	-	2741	2892	2117	4237	28	-
Polypedilum nubeculosum gr.	159	72	4	-	-	-	173	83	2	-	-	-
Dicrotendipes nervosus gr.	317	266	8	-	-	-	143	90	1	-	-	-
Demicryptochironomus	8	-	-	-	-	-	2	-	-	-	-	-
Cladopelma lateralis gr.	16	60	-	-	-	-	3	2	-	-	-	-
Paracladopelma camptolabis gr.	-	4	12	-	-	-	-	0	3	-	-	-
Tanytarsus	8	-	4	-	-	-	1	-	4	-	-	-
Chironomidae total	2726	2512	1420	992	155	16	3636	4280	3167	4709	175	42
%	64	73	98	96	75	9	26	90	100	99	72	7
CERATOPOGONIDAE	32	32	-	-	-	-	9	10	-	-	-	-
Trichoptera												
Ecnomus tenellus	-	4	_	-	-	-	-	11	-	-	-	-
Mystacides azurea	8	4	4	-	-	-	9	2	5	-	-	-
Other insecta total	40	40	4	-	-	-	18	23	5	-	_	_
%	40	1	4 0	-	-	-	0	1	0	-	-	-
GENERAL TOTAL	4277	3453	1452	1032	207	179	14180	4751	3179	4760	244	589

Tab. 5. Annual mean values for density and biomass, recorded in 1994, for different taxonomic entities in the various stations. Values relating to Chironomidae include larvae and pupae. Percentages are also presented.

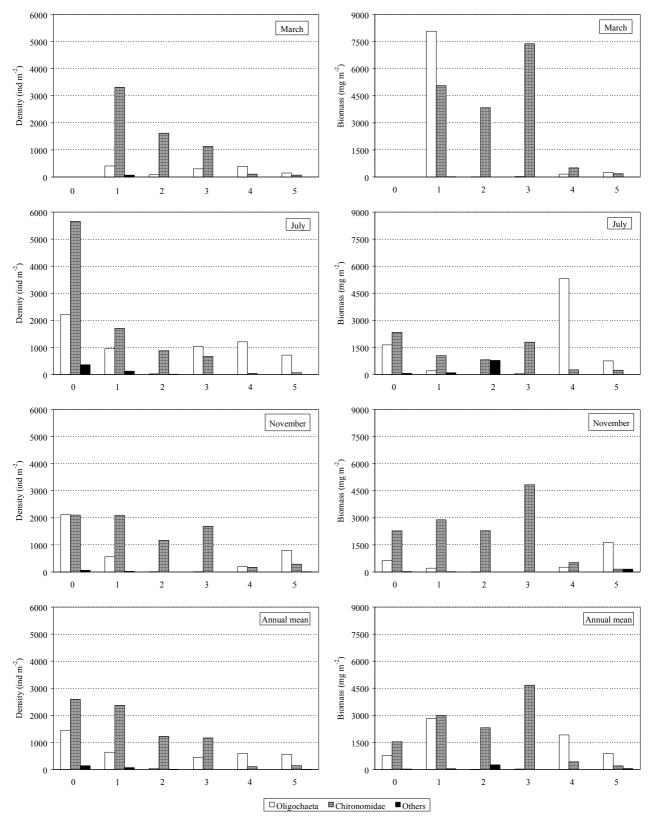


Fig. 5. Seasonal variations and annual means of density and biomass for Oligochaeta, Diptera Chironomidae and other organisms in the different stations considered in 1993.

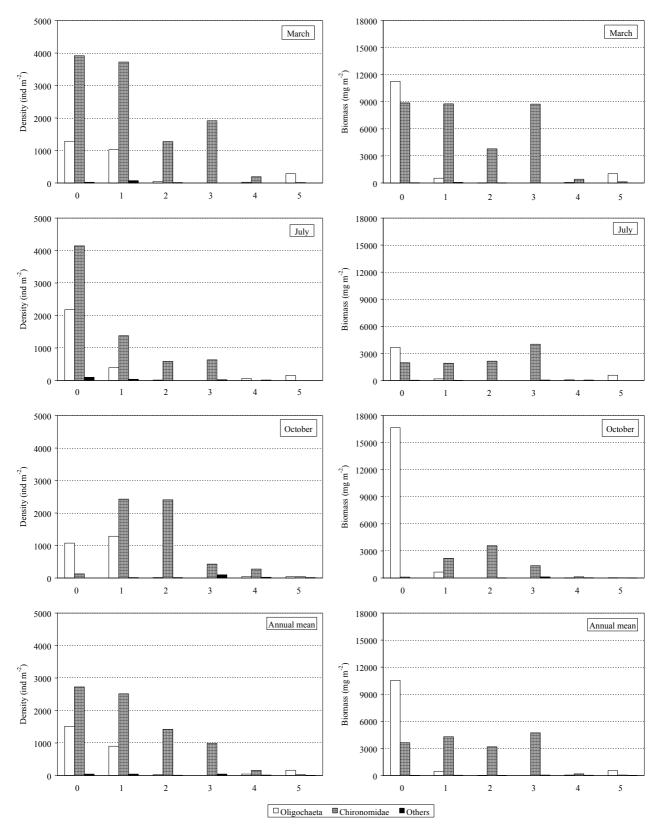


Fig. 6. Seasonal variations and annual means of density and biomass for Oligochaeta, Diptera Chironomidae and other organisms in the different stations considered in 1994.

4. CONCLUSIONS

The results of this investigation seem to indicate that liming the lake has facilitated the repopulation of the Buccione Basin by macrobenthic fauna, at least in numerical terms. As has been observed in other acidified lakes which have undergone neutralisation treatments (Eriksson *et al.* 1983; Raddum *et al.* 1986; Keller *et al.* 1990; Kelly 1991), this basin was colonised initially by some groups of organisms such as Oligochaeta and Diptera Chironomidae, which showed richer populations in the littoral layers.

The qualitative reconstruction of the macrobenthos has however been slower than that of other biological communities such as phyto- and zooplankton, which began to reorganise rapidly in the earliest stages of the lake's recovery (Morabito *et al.* 2001; Bonacina & Pasteris 2001). In fact, in 1993-94, around three years after liming, the Oligochaeta community in the Buccione Basin was still made up only of a few Naididae and Tubificidae species varying in their distribution according to depth. Tubificidae species typical of non-acidified lakes, such as *Spirosperma ferox*, which were present in Lake Orta before it was polluted, were still absent (Bonacina *et al.* 1986).

The bathymetric distribution of the two most numerous Tubificidae species, *Limnodrilus hoffmeisteri* and *Tubifex tubifex*, is especially noteworthy. The former has greatly increased in numbers in the littoral zone, where, being sensitive to acidification (Wiederholm & Eriksson 1977; Merilainen & Hynynen 1990), the Tubificidae were poorly represented in the years prior to the liming (Oioli 1969), while the latter mainly colonised the deep layers. It should also be remembered that up to 1996-97 *Tubifex tubifex* comprised almost all the macrobenthos in the entire deep zone of the lake (Baudo *et al.* 2001), where it was already present at the beginning of the 80s when the lake was strongly acidified (Bonacina & Bonomi 1984).

The different bathymetric distribution of the two species of Tubificidae is dependent on a set of factors among which those linked to the chemical-physical characteristics of the sediments and the depth undoubtedly play a major role (Nalepa & Thomas 1976; Barton 1988; Mc Murtry *et al.* 1983). Both species prefer a substrate made up of silt and sand, but while *Limnodri-lus hoffmeisteri* in lakes normally colonises the littoral and sub-littoral layers, *Tubifex tubifex* has a preference for the deep zone, thanks to its ability to reproduce and increase even at temperatures below 10 °C and to survive for long periods at low oxygen concentrations (Milbrink 1980; Hoffmann *et al.* 1986; Anlauf & Neumann 1997).

Furthermore, *Tubifex tubifex* is a species which can adapt to environments with different degrees of trophy and pollution, and it is undoubtedly more tolerant of acidification than other Tubificidae species (Wiederholm & Dave 1989; Johnson *et al.* 1992). In Lake Orta

it has also shown great resistance to the high content of heavy metal in the sediment, particularly copper, chrome, zinc and nickel; in the past this derived from industrial waste, and since the liming it has gone on increasing, especially in the upper layers, with the precipitation of these elements from the water column (Baudo & Beltrami 2001). The high degree of ecological adaptability of *Tubifex tubifex* is undoubtedly helped by its genetic variability: within the species, in fact, various ecological races correlated with different environmental conditions have been identified (Anlauf & Neumann 1997).

The Diptera Chironomidae community is still also rather poor in quality. After the liming, some larval forms of Orthocladiinae in particular increased in number, especially Parakiefferiella bathophila and some genera of Tanypodinae (Procladius and Ablabesmyia) and Chironominae (Dicrotendipes and Chironomus thummi gr.), which were already present in the littoral zone of the lake during its phases of pollution. The number of larval forms found along the transect examined is still markedly lower than the number which would normally characterise this type of community in natural lake environments. The Tanytarsini, for instance, found at depths above 4 m are still very scarce (De Margaritis 1995); they were well represented before the lake was polluted, as is shown by the analysis of a sediment core taken in the Buccione Basin in 1993 at a depth of 22 m (Nocentini, pers. comm.). It is noteworthy that in 1996-97 Oligochaeta and Chironomidae communities composed of an extremely small number of systematic entities were still present not only in the Buccione Basin, but also in the other basins of the lake, particularly those of Pettenasco and Omegna (Baudo et al. 2001).

The slowness with which, several years after the liming, the lake has been re-colonised by many systematic entities of Oligochaeta and Chironomidae is probably due to the marked increase in metal concentrations in the sediment, produced by the neutralisation of the water. It is also likely that changes in the lake water chemistry, especially in pH, which very soon after the recovery intervention was showing values around or higher than 6, encouraged the liberation in the water and the sediment of metals in ionic forms, which are more easily accumulated in benthic organisms through their body membranes or from food sources, but which are highly toxic (Campbell & Stokes 1985; Gerhardt 1993).

Food is an important channel for the transfer of toxins from the surrounding environment to bottom organisms like Oligochaeta and Chironomidae, which feed largely on algae, detritus, and associated microorganisms, which have sedimented to the bottom or are suspended in the water layers immediately above the bottom (Brinkhurst 1971; Berg 1995). In Lake Orta these particles are mostly autochthonous in origin, and come mainly from particulate matter produced in the water column and sedimented.

The toxins accumulating in benthic organisms may have a negative effect on their fecundity and their lifecycle, and may reduce markedly the numbers of systematic entities in the community (Wiederholm & Dave 1989; Johnson et al. 1992). A drop in the number of Chironomidae taxa has for example been recorded frequently in environments contaminated by copper, chrome and zinc, like Lake Orta (St. Louis 1993; Lindegaard 1995). However, it should be stressed that the effects of these toxins on the organisms depend on the type of metal available, and on abiotic factors like pH and temperature, as well as on biotic factors like size and body weight, or stage of development (Chapman et al. 1980; Gerhardt 1993; Lindegaard 1995). Nevertheless, there are populations that are able to survive in the presence of high metal concentrations, because they possess efficient detoxification mechanisms, or can lower the amount of metal accumulated in their bodies either through excretion or, as in the case of insects, through moulting and emergence. Populations which live for generations in contaminated environments can also develop a genetic resistance to metals (Gerhardt 1993; Lindegaard 1995).

In the Buccione Basin, the macrobenthic fauna is still qualitatively poor in the sub-littoral and deep layers, where some species, particularly the more tolerant forms such as *Tubifex tubifex*, *Chironomus thummi* gr. and *Procladius*, are present in considerable quantities. It will probably require some years before these zones become re-populated with systematic entities characteristic of environments with low pollution, particularly other Tubificidae species. It should be remembered that these species live several centimetres below the surface of the sediment, where the values of metal concentrations may, as in Lake Orta, be higher than those of the surface layers (Beltrami *et al.* 1999).

Interestingly in this context, some Tubificidae species were found in 1996-97 in the Tortirogno Basin (Baudo et al. 2001); compared to the other basins of the lake, this one has better environmental conditions, mainly because of low values of metals in the sediment. At a depth of 25 m, besides Limnodrilus hoffmeisteri and Tubifex tubifex, the presence of Branchiura sowerbyi and Aulodrilus pluriseta was discovered. In the same basin in 1997-98, significant quantities of Aulodrilus pluriseta were observed between 8 and 10 m (Baudo, pers. comm.). At the same depth abundant populations of larval forms were recorded among the Chironomidae. They belong to the genera Cladotanytarsus and Tanytarsus among Tanytarsini, and to some genera of Chironomini, such as Pagastiella orophila, Cladopelma gr. lateralis, Paracladopelma gr. camptolabis, commonly found in the littoral-sublittoral zones of subalpine lakes.

A qualitatively more varied macrobenthos has been recorded in the other basins of the lake, especially in the shallow layers. In 1993-94, a higher number of Oligochaeta species, in particular Naididae, and larval forms of Chironomidae, was found in the littoral layers of the Buccione Basin than in the deeper layers. Also present in the community were Trichoptera, with the species *Mystacides azurea* and *Ecnomus tenellus*, Megaloptera, with the genus *Sialis*, and Diptera Ceratopogonidae, which were observed in the lake before liming.

Between 1993 and 1998 (De Margaritis 1995; Bielli & Tesauro 2001), representatives of other taxonomic groups were found around the perimeter of the lake at depths above 4 m. Of particular note among these are the Oligochaeta Lumbriculidae and Enchytraeidae, and the Insecta Odonata, Ephemeroptera, Plecoptera and Coleoptera. Also present were Turbellaria, Hydracarina, and Crustacea (*Asellus aquaticus*).

The structure of the littoral and the profundal macrobenthos will undoubtedly undergo further changes in the next few years, but the paucity of earlier data on the deep-water fauna will make it difficult to compare the future situation of the community with the situation existing prior to the pollution of the lake. Nevertheless, it is unlikely that a stable and qualitatively well-organised community will be established, particularly in the deep layers, as long as there is no reduction in metal concentrations at sediment level. All the same, the future evolution of the macrobenthos should be monitored, also because the lake's trophic state will probably increase as a consequence of large inputs of phosphorus from the watershed. These could cause extensive changes in the physical-chemical and biological parameters of the water and the sediments, with major implications for the colonisation of benthic organisms.

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